### **RESEARCH ARTICLE**

**Open Access** 

# Spatial distribution and seasonal movement patterns of reintroduced Chinese giant salamanders



Lu Zhang<sup>1,2,3</sup>, Hu Zhao<sup>4</sup>, Scott Willard<sup>3</sup>, Qijun Wang<sup>4</sup>, Wei Jiang<sup>4</sup>, Hong-Xing Zhang<sup>4</sup> and Andrew Kouba<sup>2,5,6\*</sup>

#### **Abstract**

**Background:** Very little is known about the temporal or spatial movement patterns of Chinese giant salamanders (*Andrias davidianus*) due to their rarity, remote habitat and secretive nature. Commercial breeding farms provide a unique opportunity as a source of animals for reintroduction and spatial ecology studies, which will help inform conservation management efforts for this threatened species. We surgically implanted radio transmitters into the body cavity of 31 juvenile giant salamanders, and these salamanders were subsequently released into two small river systems (Donghe and Heihe Rivers) located in the Qinling Mountains of central China and were monitored daily from May 2013 to August 2014.

**Results:** Only two salamanders survived through the end of the project at the Heihe River compared with 12 at the Donghe River, thus movement data for salamanders released at the Heihe river are described individually. The overall sedentariness (ratio of no movement to all observations) for the two salamanders at the Heihe River was 0.29 and 0.28 compared to the average sedentariness of  $0.26 \pm 0.01$  for the 12 salamanders at the Donghe River. Mean daily movement was  $15.4 \,\mathrm{m} \pm 0.7$  at the Heihe River compared to  $9.3 \,\mathrm{m} \pm 0.3$  at the Donghe River. Overall linear home range (LHR) was 246 m and 392 m for the two salamanders at the Heihe River, compared with a mean LHR of  $227.2 \,\mathrm{m} \pm 70.5$  at the Donghe River. The Donghe salamanders exhibited different movement patterns across seasons, having higher sedentariness, shorter daily movement, and smaller LHR in winter than in summer. Upstream dispersal and fidelity to release site were recorded at both rivers. The mean dispersal distance for the Donghe River salamanders was  $145.3 \,\mathrm{m} \pm 61.9$ , while the two surviving salamanders at the Heihe River had a dispersal distance of 211 m and 205 m.

**Conclusions:** This project provides important insights on the movement ecology of a large aquatic salamander species, and in particular, our results may assist with reintroduction efforts by developing best management practices on when and where to release animals as a conservation strategy.

**Keywords:** Andrias davidianus, Aquatic salamander, Daily movement, Dispersal, Home range, Radio-telemetry, Reintroduction

#### **Background**

Around the world, amphibian declines and local extinctions are accelerating at an alarming rate, with >40% of all amphibians listed as threatened on the IUCN Red List [1]. This percentage is higher than in birds (14%)

and mammals (25%), raising global concerns about the loss of amphibian biodiversity. The various factors contributing to the decline of amphibians include habitat loss, over-utilization, disease, pollution, and climate change [2–5]. Among amphibians, species with a larger body mass or that live in aquatic habitats may have a higher risk of decline [6, 7]. Both risk factors are exemplified by species in the Cryptobranchidae family, which have the largest body size among living amphibians and exhibit totally aquatic life histories. There are three cryptobranchid species including the hellbenders (*Cryptobranchus* 

<sup>&</sup>lt;sup>5</sup>Conservation and Research Department, Memphis Zoo, 2000 Prentiss Place, Memphis, TN 38112, USA





<sup>\*</sup> Correspondence: a.kouba@msstate.edu

<sup>&</sup>lt;sup>2</sup>Department of Wildlife, Fisheries and Aquaculture, Mississippi State University, Starkville, MS 39762, USA

Zhang et al. BMC Zoology (2019) 4:7 Page 2 of 11

alleganiensis) in North America, the Japanese giant salamanders (JGS, Andrias japonicus) and Chinese giant salamanders (CGS, Andrias davidianus) in Asia. According to the IUCN Red List, the hellbender and the JGS are listed as Near Threatened, while the CGS as Critically Endangered [8–10]. Furthermore, the two subspecies of hellbender are listed differently under the US federal Endangered Species Act (ESA), with C. a. bishopi classified as Endangered, and C. a. alleganiensis as Threatened. Conservation efforts are needed for these giant salamanders, especially for the Critically Endangered CGS.

Translocation and reintroduction are increasingly implemented in restoration efforts of threatened amphibian species [11]. Thus, having an understanding of the target species' movement ecology is crucial to designing reintroduction projects; for example, with aquatic species, choosing a long enough stretch of river for released animals to diversify is important for the establishment of a reintroduced population. These three giant salamander species are reported to be highly sedentary [12–14]. This trait may be beneficial for establishing populations in the wild, as moving away from release site is one of the most common reported causes to reintroduction failure [11], and penning prior to release would reduce post-release dispersal thus contribute to population establishment [15, 16]. Unfortunately, very little is known regarding other aspects of movement for CGS and JGS, although there are a number of studies on movement ecology for hellbenders [14, 17-19] that can be used for comparison.

As an endemic species, CGS were once found widely in the mountain tributaries of the three major river systems-Pearl, Yellow and Yangtze Rivers-in central and southern China [20]. However, since the 1950s CGS populations have declined dramatically, largely due to over-exploitation for human consumption, water pollution, and habitat destruction such that they are now almost extinct in the wild [10, 21-23]. Unlike hellbenders, no published studies have radio tracked wild CGS and their spatial ecology remains largely unknown. The only available telemetry study monitored 4 captivereared CGS (two adult males and two adult females) reintroduced to the wild over a 6 month period [13]. This study found that all 4 of their released animals moved upstream, settled in one location within 10 days post-release and had an average home range of 34.75 m<sup>2</sup> [13]. Although this initial publication studying CGS movement revealed some interesting aspects on the species movement and habitat selection, more animals tracked over a longer period of time would be a valuable addition to understanding the species' movement ecology across seasons.

Due to CGS's rarity and associated challenges with telemetry approval for wild animals in China, captive populations may be a better sink for studying movement, especially given these populations will be used for future reintroduction efforts. Critical information on movement ecology under a controlled design is possible with captive-bred individuals and has been demonstrated in the high profile recovery efforts of the California condor (Gymnogyps californianus, [24]). Although captive breeding for the purpose of CGS conservation has not been reported, commercial breeding in Andrias farms has been highly prosperous over the past decade in China [25]. In these farms, hundreds of thousands of animals each year are produced for human consumption, as they are considered a delicacy in the Chinese food market. The farms might also provide animals for restoration efforts, which relieves pressure on wild populations at risk of poaching. The release of farm-bred animals also provides an ideal opportunity to examine CGS movement ecology that can be used to optimize reintroduction projects.

We reintroduced 31 juvenile CGSs at two sites in the Qinling Mountains in central China and monitored their movements through radio-telemetry. We have previously reported survivorship [26] and habitat use [27] but describe here the post-release movement of the salamanders over an extended period of 16 months, from May 2013 to August 2014. Our specific objectives were to: 1) assess activity patterns including sedentariness (ratio of no movement to all observations) and daily movement; 2) determine home range sizes and overlap of home ranges with conspecifics; 3) compare seasonal movement patterns; and 4) assess dispersal distances and variables that may affect dispersal patterns. We hypothesized that the released CGSs would make short-distance movements of a comparable magnitude to the movements observed by hellbenders [18, 19], and have an up-stream biased dispersal similar to other aquatic salamanders [28, 29]. Results from this study will help to fill in knowledge gaps in movement ecology of CGS, and may provide suggestions for future reintroduction projects on better monitoring design.

#### Results

#### Sedentariness and daily movement

A total of 5939 salamander locations were recorded during the study, of which 4441 were from the Donghe River. The overall sedentariness ranged from 0.09 to 0.53 at the Heihe River, and 0.18 to 0.32 at the Donghe River (Table 1). Only two salamanders in the Heihe cohort survived at least four seasons, and for each an overall sedentariness of 0.29 and 0.28 was determined. At the Donghe River, 12 salamanders survived at least four seasons, with an average overall sedentariness of 0.26  $\pm$  0.01. The mean daily movement was 15.4  $\pm$  0.7 m (n = 979, 4–298 m) at the Heihe River, and 9.3  $\pm$  0.3 m (n = 3185, 4–880 m) at the Donghe River. Although our

Zhang et al. BMC Zoology (2019) 4:7 Page 3 of 11

**Table 1** Parameters describing the post-release movement and home ranges of individual Chinese giant salamanders reintroduced to the Heihe (3 year old animals) and Donghe (5 year old animals) rivers in central China

Animal ID	n <sup>a</sup>	Sedentariness	Daily movement (m) <sup>b</sup>	LHR (m) <sup>c</sup>	MCP (m <sup>2</sup> ) <sup>d</sup>	Status <sup>€</sup>
Heihe						
202	2	-	17 (n = 1)	10	-	$M^*$
235	403	0.29	9 (n = 274) [4–107]	246	18,876	А
251	47	0.29	12 ( <i>n</i> = 29) [4–100]	194	5534	M*
273 <sup>f</sup>	126	0.38	9 (n = 73) [4–262]	2364	302,433	U
291	4	-	6 (n = 3) [4–77]	80	210	M <sup>-</sup>
351	5	-	10.5 (n = 2) [9–12]	16	78	$M^*$
392	12	0.09	10.5 ( <i>n</i> = 10) [5–89]	176	2588	$M^*$
412	180	0.22	7 (n = 131) [4–70]	64	3571	U
431 <sup>f</sup>	121	0.25	10 ( <i>n</i> = 84) [4–132]	1083	24,726,132	U
472 <sup>f</sup>	46	0.53	20 ( <i>n</i> = 19) [4–85]	9524	959,773	U
531	397	0.28	8 ( <i>n</i> = 271) [4–298]	392	25,625	Α
571	51	0.30	11 ( <i>n</i> = 30) [4–60]	91	4301	U
610	46	0.40	6 (n = 24) [4–215]	401	17,182	U
710	3	-	7 (n = 2) [6–8]	12	1	$M^*$
730 <sup>f</sup>	55	0.41	10.5 (n = 26) [4–113]	3865	574,909	U
Donghe						
101	385	0.29	6 (n = 267) [4–64]	389	11,149	А
211	326	0.31	6 ( <i>n</i> = 216) [4–77]	156	3868	А
312	155	0.32	6 (n = 98) [4–44]	118	2630	M <b>-</b>
332	332	0.20	8 ( <i>n</i> = 257) [4–32]	46	1156	А
371	339	0.24	6 (n = 247) [4–49]	76	2086	А
450	364	0.22	7 (n = 274) [4–34]	41	828	А
490 <sup>f</sup>	327	0.27	6.5 (n = 225) [4–48]	722	44,955	А
511	401	0.19	7 (n = 316) [4–37]	84	1733	А
550	76	0.32	6 ( <i>n</i> = 44) [4–93]	261	6099	M <sup>-</sup>
591 <sup>f</sup>	262	0.18	9 ( <i>n</i> = 207) [4–880]	1730	159,357	U
630 <sup>f</sup>	332	0.29	6 ( <i>n</i> = 217) [4–21]	762	52,728	U
651	59	0.27	7 (n = 36) [4–30]	41	634	U
671	335	0.21	8 ( <i>n</i> = 254) [4–54]	151	3610	А
751	345	0.22	7 (n = 254) [4–100]	560	20,390	А
770	77	0.29	8 ( <i>n</i> = 45) [4–18]	93	1227	M <sup>=</sup>
790	326	0.30	7 (n = 221) [4–79]	542	22,367	А

<sup>&</sup>lt;sup>a</sup>Total number of observations collected

data indicated that salamanders can move large distances (up to 880 m), their mean daily movement was relatively short, with 74% of daily movements  $\leq 10$  m at the Donghe River and 58% at the Heihe River.

#### Home range and overlap of home ranges

Linear home range (LHR) during the entire study period differed considerably among salamanders, ranging from 10 to 9524 m at the Heihe River, and 41 to 1730 m at

<sup>&</sup>lt;sup>b</sup>Median of distances between sequential daily locations that were > 3 m. Numbers in parentheses indicate sample size and numbers within brackets represent the range of observed values (minimum cut-off of 4 m to represent a daily movement)

<sup>&</sup>lt;sup>c</sup>Linear Home Range: length of river section occupied by the salamander during the period it was monitored

<sup>&</sup>lt;sup>d</sup>Minimum convex polygons

eSalamander status: A = Alive, U=Undetermined, M = Mortality, with superscripts explaining causes of mortality, where \* = dehisced sutures,

<sup>◆</sup> = flood,  $\blacksquare$  = unknown

fSalamanders moved downstream during floods

Zhang et al. BMC Zoology (2019) 4:7 Page 4 of 11

the Donghe River (Table 1). Minimum convex polygons (MCP) showed the same variations. Not surprisingly, the largest home ranges were from salamanders that moved downstream during floods, while the smallest home ranges were from salamanders that died soon after release. Excluding salamanders that were swept downstream during floods and those that did not survive at least four seasons, each of the two salamanders surviving at the Heihe River had an overall LHR of 246 m and 392 m, and an overall MCP of  $18,876 \,\mathrm{m}^2$  and  $25,625 \,\mathrm{m}^2$ . The Donghe salamanders had a mean LHR of  $227.2 \pm 70.5 \,\mathrm{m}$  (n = 9, range:  $41-560 \,\mathrm{m}$ ), and a mean MCP of  $7465.2 \pm 2828.8 \,\mathrm{m}^2$  (n = 9, range:  $828-22,367 \,\mathrm{m}^2$ ). LHRs of these  $11 \,\mathrm{salamanders}$  were strongly correlated with their MCPs (*Spearman rho* = 0.945, p < 0.001).

For each salamander at the Donghe River, the number of neighbors with a weekly overlapping LHR decreased quickly post-release and remained relatively stable in the following weeks (Fig. 1a). Similarly, the percentage of overlap also rapidly decreased in the first 10 weeks and then became relatively stable, with an average overlap of 39.4% (range: 0–99.1%, Fig. 1b). After the 50th week, we started to lose signal due to battery depletion, thus the 0 overlap in Weeks 50–52 might be biased (i.e., the overlapping neighbor's data may be missing because we could not locate it anymore, when in fact it could have remained in the river) and were presented here for reference only. Despite this general overlap of home ranges, we never found two salamanders under a single boulder simultaneously.

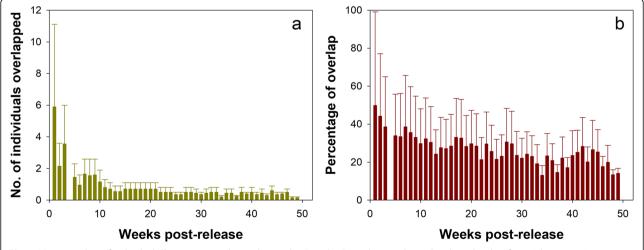
## Seasonal variation in movements of the Donghe salamanders

Salamanders at the Donghe River had different sedentariness (Fig. 2a), mean daily movement (Fig. 2b), and

LHR among seasons (Fig. 2c). Specifically, salamanders had a higher sedentariness (*Friedman chi-squared* = 11.045, p = 0.026), shorter daily movement (*Friedman chi-squared* = 12.447, p = 0.014), and shorter LHR (*Friedman chi-squared* = 9.542, p = 0.049) in winter than in summer. Measurements in spring and autumn were usually between those of winter and summer and did not differ significantly. The only exception was sedentariness in summer 2013 (Fig. 2a), which was higher than sedentariness in summer 2014 (p = 0.007) and did not differ from that in winter 2014 (p = 0.939).

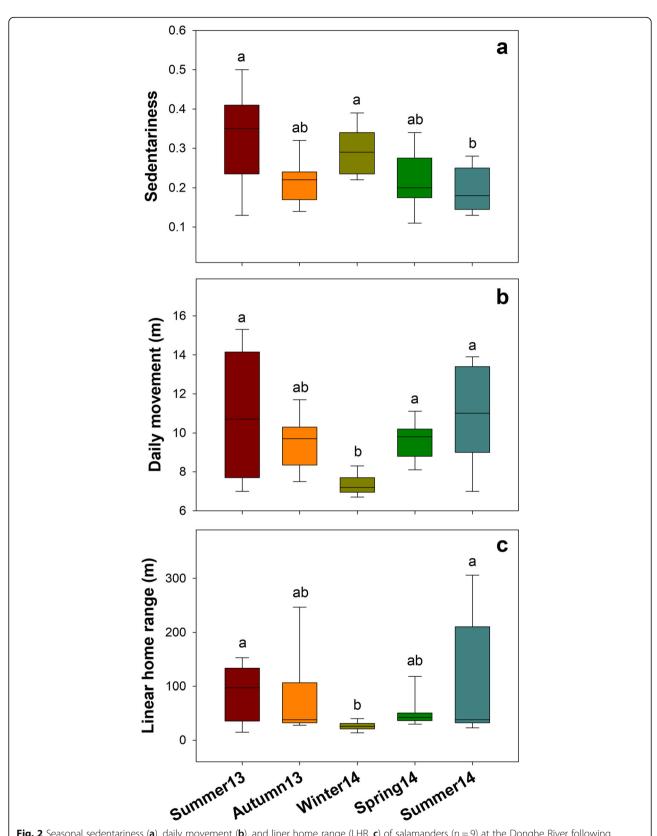
#### Dispersal patterns

Released salamanders showed all three types of dispersal patterns - moving upstream, moving downstream, and fidelity to release site (Figs. 3, 4). All downstream dispersals were associated with flood events. Among salamanders with different types of dispersal at both rivers, no difference was found on their initial body mass (Donghe: H = 1.000, p = 0.607; Heihe: H = 2.786, p =0.248) or body condition (Donghe: H = 1.250, p = 0.535; Heihe: H = 1.071, p = 0.585). Regardless of direction, the dispersal distance (distance from release site to the centroid of the last weekly LHR) of salamanders ranged from 1 to 572 m at the Donghe River (Fig. 3a-d) and 1 to 6130 m at the Heihe River (Fig. 4 a-d). Similar to home ranges, the largest dispersal distances were from salamanders that moved downstream during floods, while the smallest dispersal distance was from salamanders that died or disappeared at an early stage postrelease. Excluding salamanders that moved downstream during floods and those that did not survive at least four seasons, the mean dispersal distance was  $145.3 \pm 61.9$  m (n = 8, range: 4-451 m) at the Donghe River, and 211 m



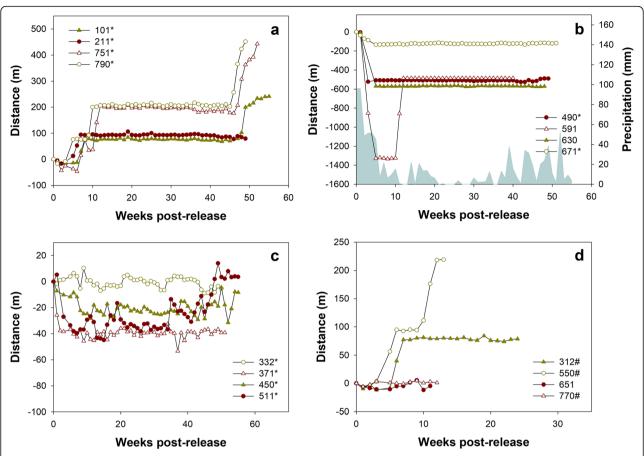
**Fig. 1** Mean number of individual Chinese giant salamanders at the Donghe River that overlapped with each other from July 2013 – August 2014 (a), as well as the mean percentage of overlap (b). From the 50th week forward we started to lose radio signal due to battery depletion, thus 0 overlap in Weeks 50–52 may be biased and are only presented for reference

Zhang et al. BMC Zoology (2019) 4:7 Page 5 of 11



**Fig. 2** Seasonal sedentariness (**a**), daily movement (**b**), and liner home range (LHR, **c**) of salamanders (n = 9) at the Donghe River following release. Seasons with the same lowercase letters indicate there is no difference between the two. The upper whiskers represent the 90th of the percentile while the lower whiskers represent the 10th of the percentile of the data

Zhang et al. BMC Zoology (2019) 4:7 Page 6 of 11



**Fig. 3** Dispersal distance from the release site for every Chinese giant salamander reintroduced at the Donghe River (July 2013 – August 2014), including those that moved upstream (**a**), downstream (**b**), showed fidelity to their release sites (**c**), and those that died shortly post-release (**d**). Salamander identities were listed in figure legends, with \* indicating the animal was confirmed alive at the end of the project, and # indicating the animal was confirmed dead. Shaded area in (**b**) indicates the amount of weekly precipitation in millimeter

and  $205\,\mathrm{m}$  for the two surviving salamanders at the Heihe River.

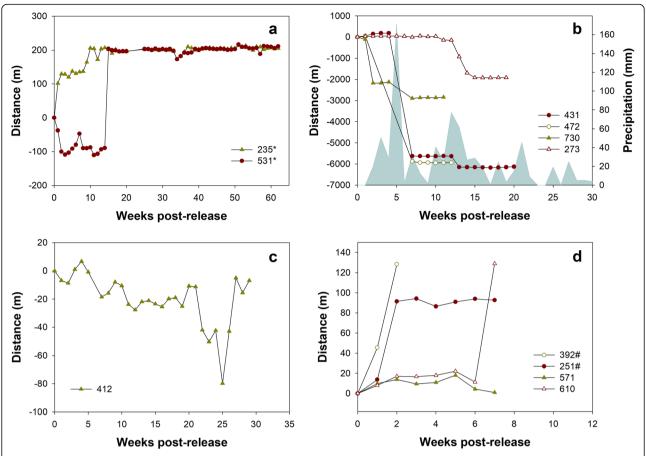
The 4 up-stream dispersing salamanders at the Donghe River found a suitable refuge within 12 weeks and remained at those sites for almost an entire year. Three of the four salamanders displayed a second round of movement and dispersal between the 46th–48th week post-release (Fig. 3a). Upstream dispersal of these three animals continued until the transmitters ran out of battery power.

#### **Discussion**

Although giant salamanders are generally considered as highly sedentary and having limited mobility, our study revealed that reintroduced CGSs moved quite frequently throughout the year. The year-round sedentariness for our salamanders was under 0.31, indicating these animals moved  $\sim 70\%$  of the days during our study period. The previous telemetry study on this species [13] does not report an overall sedentariness for their salamanders

during their monitoring period. Instead, they indicate that their animals moved upstream within 3-10 days post-release and had a second round of upstream movement after their dens were destroyed by a flood. Thus, we could not directly compare our results with their findings and make meaningful predictions. However, there are several studies on the movement ecology of hellbenders, which we can compare our study against. Burgmeier et al. [14] indicated that wild hellbenders moved during ~ 25% of their observations, and Bodinof et al. [18] reported captive-reared hellbenders moved < 50% of their observations post-release. Compared to these results, we found that CGSs moved more frequently than the hellbenders. Lacking movement data from wild CGSs, we cannot determine whether the high frequency of movement of our animals was because they were captivereared, continuously seeking a suitable refuge or if it was a biological trait of this species. More studies on the movement of captive-reared vs. wild aquatic salamander species, may help to answer this specific question.

Zhang et al. BMC Zoology (2019) 4:7 Page 7 of 11



**Fig. 4** Dispersal distance from the release site for every Chinese giant salamander reintroduced at the Heihe River (May 2013 – July 2014), including those that moved upstream (**a**), downstream (**b**), showed fidelity to their release sites (**c**), and those died shortly post-release (**d**). Salamander identities were listed in figure legends, with \* indicating the animal was confirmed alive at the end of the project, and # indicating the animal was confirmed dead. Shaded area in (**b**) indicates the amount of weekly precipitation in millimeter

We found that reintroduced CGSs generally move short distances, with a mean daily movement of 9–15 m, although they were able to move long distances, up to 880 m within a single day. These results were comparable to reports on hellbenders, which found that hellbenders travel on average 3 m daily [18], 4 m weekly [19] or 28 m for less regular movements [14]. Similarly, long-distance movements have been reported on hellbenders with a range from 347 to 990 m [14, 17, 18]. Our findings were in line with conclusions from studies on hellbenders [14, 18] and other amphibians [30, 31], in that we found giant salamanders are capable of moving relatively long-distance over a short period of time, but that their tendency is to limit their movements to short distances.

The average overall LHR for the Donghe salamanders was  $227 \pm 70$  m (range: 41-560 m), while those of the two salamanders that survived at the Heihe River were 246 m and 392 m. These measurements were comparable to results from the previous study on CGSs [13],

which indicated that the distance from salamanders' initial locations to their settlement locations were 204-554 m, but probably larger than those reported for hellbenders, e.g.,  $144 \pm 58 \,\mathrm{m}$  [14] and  $23-110 \,\mathrm{m}$  [32]. When comparing MCPs, salamanders in our study also had larger home ranges  $(7465.2 \pm 2828.8 \text{ m}^2)$  than hellbenders  $(2211.9 \pm 990.3 \,\mathrm{m}^2, [14])$ . However, these home ranges were still relatively small considering that salamanders can move up to 880 m within a single day. Together, with results of sedentariness and daily movement, we may infer that reintroduced CGSs made frequent shortdistance and non-directional movements, post-release. We may conclude that similar to hellbenders [18], reintroduced CGSs had a relatively high site fidelity to their release site, at least in their first year in the wild. These traits may be beneficial for the establishment of new populations in the wild [11, 16, 33]. However, these traits, together with the small initial population size, may result in limited gene flow and reduced effective population size, which impede the persistence of Zhang et al. BMC Zoology (2019) 4:7 Page 8 of 11

populations. Therefore, several rounds of reintroduction efforts may be needed before a self-sustaining population could be established in the wild.

The overlapping LHR of the Donghe salamanders for most of the study period indicated that reintroduced juvenile salamanders had a high tolerance of other individuals, at least in their first year in the wild. This tolerance and overlap of home ranges may reflect that they were raised in larger groups in captivity, in accordance with predictions from the mechanism called natal habitat preference induction (NHPI, [34]). However, this mechanism remains to be further tested since NHPI was not found to be supported on several pond-breeding salamanders [35]. Burgmeier et al. [14] reported that overlapping home ranges of adult hellbenders only occur in summer, and they had once located a male and a female under the same shelter during the breeding season. However, in our study salamanders had overlapping home ranges throughout all four seasons in the wild. One possible explanation for this overlap was that our salamanders were still juveniles and juveniles may not be as territorial as adults. Moreover, because they were juveniles, we never found two or more giant salamanders under the same shelter due to breeding. Hence, for monitoring and management convenience, future reintroduction projects could release juvenile CGSs relatively close to each other at the beginning (in rivers with plenty of food resources) without significant injury or mortality due to territorial battles.

We found that salamanders moved less frequently and shorter distances in winter than in summer, which is similar to the movement observed by hellbenders [14]. As an ectotherm, this movement pattern may largely be related to temperature since CGSs are found to increase their metabolic rate and food intake as temperature increases, as long as the temperature is below their upper tolerance limits [36]. Many amphibian and reptile species hibernate during winter to cope with low temperatures (e.g., [37, 38]). Salamanders in our study showed reduced movement (both frequency and distance) in winter; however, they did not seem to hibernate during this cold period as their average sedentariness was still around 0.3 in winter. Interestingly, salamanders had a high sedentariness in their first summer in the wild, as high as that in winter, which was much higher than in the second summer. Reintroduced/translocated animals may exhibit exploratory movement after release, and would move more frequently with longer distances to search for suitable habitat (e.g., [39-41]). Our observed movement patterns were contrary to these findings. Salamanders moved less frequently in the first summer than in the second summer, although when they did move, they moved long distances as seen in the second summer. This may be explained by an alternative theory, that naive animals reintroduced to a novel environment usually move very little shortly after release because their fear of predators would reduce their mobility to avoid predator encounter rates [42]. Yet, when they did move, their moving ability was more affected by seasons (temperature). Additional studies are needed before we can make any accurate predictions related to this movement pattern.

In contrast to our hypothesis that salamanders would primarily disperse upstream, we found almost an equal number of salamanders showed fidelity to release site as upstream dispersal (6 vs. 6 at Donghe, 1 vs. 2 at Heihe, respectively). Furthermore, 4 animals were pushed downstream during floods at each river. We did not find any difference based on body mass or body condition between salamanders with different dispersal patterns, probably due to a relative equivalent body condition of individuals within the same age grous from the same breeding farm. We propose that the age of the animals may have an effect on their dispersal directions. For example, Bodinof et al. [18] reported that 20 out of 26 juvenile hellbenders dispersed downstream post-release, whereas Cecala et al. [29] found that large larval salamanders exhibit upstream movement, but not small larvae. Zheng and Wang [13] reported that all four of their older study animals moved upstream and none of them were pushed downstream by floods, although floods destroyed their dens and triggered a second round of upstream movement. The four salamanders in Zheng and Wang's study were adults and may have been more resilient to floods than the juveniles in our study. In support of this proposal, the Heihe salamanders, which were 2 years younger than the Donghe animals, were more affected by floods such that they were pushed further downstream and all disappeared with status undetermined. In contrast, salamanders that were pushed downstream during floods at the Donghe River went shorter distances and half of them were confirmed alive and continued to be monitored. These results provide supporting evidence that younger giant salamanders may be more affected by floods and are more likely to disperse downstream (passively). However, different characteristics of the rivers in our study may confound this explanation. For example, the Heihe salamanders were possibly impacted by the larger floods occurring at this river compared to Donghe river flooding (physical observation). Future studies may need to release salamanders of different ages at one site to clarify the effect of age on their dispersal patterns.

Three of the Donghe salamanders started a second round of upstream movement in June 2014. Since these animals were still juveniles (1–2 years from sexual maturity), we would not consider these movements related to breeding migration as in other amphibians [12, 43]. Rather, we propose this behavior to be dispersal of

Zhang et al. BMC Zoology (2019) 4:7 Page 9 of 11

juveniles away from their natal population [44] and the timing may be related more to seasonal changes in temperature. Unfortunately, we could not determine the individuals' sex at the time of release, and thus cannot determine whether there was a sex-biased dispersal pattern. A longer monitoring project may help to elucidate more on the dispersal behavior of reintroduced or translocated CGSs, although this would require additional transmitters be attached once battery life of the original transmitters approached its end.

#### **Conclusions**

Our study revealed that although more active than the hellbenders, captive-reared CGSs had a short-distance average daily movement, small home range, and a relatively high site fidelity to their release site, considering their ability to move long distances. These traits may contribute to the establishment of new populations in the wild. Reintroduction of this large aquatic salamander provides a good opportunity to study movement ecology of this critically endangered species in detail, which in turn may help to design more successful reintroduction projects.

#### Methods

#### Study area

Our study occurred at two headwater sites within the Oinling Mountains located in central China. The Heihe River in Zhouzhi County, Shaanxi Province (33°53'N, 108°00'E, datum = WGS84) is on the north slope of the Qinling Mountains and is a tributary in the Yellow River watershed. The Heihe River site is at an elevation of ~ 930 m and surrounded by deciduous broad-leaf forest along the river banks. The Donghe River in Ningshan County, Shaanxi Province (33°21'N, 108°16'E, datum = WGS84) is on the south slope of the Qinling Mountains and belongs to the Yangtze River watershed. This headwaters site is at an elevation of ~ 1230 m and the river banks are lined with a mixture of evergreen and deciduous broad-leaf forest. The mean width of the river stretches where we released salamanders were 15 m at the Heihe River and 9 m at the Donghe River [27]. These two sites were selected because the founder and parent stock originated in these rivers, and the wild CGSs are still found occasionally in them (according to local Fisheries Bureaus), indicating that these rivers continue to represent suitable habitat for this species.

#### Study animals

Our study involved 32 captive-reared juvenile salamanders from two age groups (3 and 5 years old; n=16 animals/age group). All weights (kg), lengths (cm) and distances (m) are reported as the mean  $\pm$  standard error (SE) of the mean. The younger cohort of animals had a

mean body mass of  $0.5 \pm 0.2$  kg and a mean total length of  $44.00 \pm 3.24$  cm; whereas, the older cohort of animals had a mean body mass of  $1.6 \pm 0.4$  kg and a total length of 63.97 ± 4.86 cm. Both cohorts were juveniles and could not be sexed. Radio transmitters (F1035, Advanced Telemetry Systems, Inc., Isanti, USA) were surgically implanted into the abdomen of all salamanders by a veterinarian on 13-16 March, 2013 [45]. One salamander from the younger cohort died before release due to postsurgical complications, hence a total of 31 salamanders were released to the wild. The younger cohort was released at the Heihe River site between 28 April - 2 May, 2013, whereas the older cohort was released at the Donghe River site on 12 July, 2013 [26]. The younger cohort was originally collected from the Heihe River as larvae by local Fisheries Bureau and head started in a commercial breeding farm. The older cohort was born in captivity; however, their parents were collected from the Donghe River. At both sites, all salamanders were released along a 50 m stretch, with a mean initial spacing of  $1.9 \pm 0.6$  m. Boulders were located throughout both rivers and each animal was released beside a rock large enough to provide shelter.

#### Radio-telemetry and data collection

We measured the body mass and total length for each salamander prior to release. Salamanders were monitored every day post-release using a radio-receiver with a 3-element Yagi antenna (R410, Advanced Telemetry Systems, Inc., Isanti, Minnesota, USA), until the depletion of the transmitter battery (radio signals disappeared gradually 16 months after activation in March 2013). Coordinates of salamander locations were recorded by handheld GPS units (60CSx, Garmin, Ltd., New Taipei City, Taiwan). We trained local field assistants at both sites to help with tracking and monitoring of released salamanders throughout the year. Salamanders were located and monitored every day except for a few days when floods occurred in the summer and early autumn (~ 15 days at the Donghe River and ~ 5 days at the Heihe River). Since boulders selected by the animals as shelter were usually too large to turn over manually, we determined the presence of salamanders using an underwater inspection camera (M12, Milwaukee Electric Tool, Brookfield, WI, USA).

#### Statistical analyses

Daily movements were calculated in ArcGIS (Version 10) as the straight-line distance (m) between locations collected across sequential days for each animal. This parameter describes the net displacement within a single day, rather than the accumulated distance a salamander moved, and is similar to many studies describing movement ecology of hellbenders (e.g., [18]). Long-distance

Zhang et al. BMC Zoology (2019) 4:7 Page 10 of 11

displacements downstream during floods were excluded from our analyses. Since the accuracy of the GPS units that we used for this project usually varied from 3 to 5 m (depending on weather condition), consecutive locations with  $\leq 3$  m distance were treated as stationary movements. Thus, sedentariness was calculated as the proportion of stationary movements ( $\leq 3$  m distances) across all daily movements observed. Mean daily movement of salamanders was calculated as the mean of all movements > 3 m.

Following Burgmeier et al. [14], we calculated both linear home range (LHR), which refers to the length of river used by an individual, and minimum convex polygons (MCP) to facilitate comparison with other giant salamander species. The overall LHR and MCP during the entire monitoring period was calculated for each salamander. We then tested the correlation between their LHRs and MCPs using Spearman's rank-order correlation test. We divided our data into weeks and generated weekly LHR for each salamander when more than four locations were recorded within a week. We determined the centroid of each weekly LHR in ArcGIS, and calculated the distance from release site to the centroids. These distances were plotted against time to show the salamanders' dispersal patterns throughout the study period. Distance to the centroid of the last weekly LHR was calculated as the dispersal distance for each salamander. Based on weekly LHR, we counted the number of individuals that overlapped with other salamanders and calculated the percentage of overlap for each week. Data from the Heihe salamanders were not used for this analysis since too many individuals died or were lost within the first couple of weeks [26].

We compared initial body mass and body condition before release for salamanders with different dispersal patterns (i.e., up-stream, down-stream, and fidelity to release site) using a non-parametric Kruskal-Wallis test. Salamanders that had a dispersal distance shorter than 100 m were categorized as fidelity to release site, whereas other salamanders were categorized as up-stream or downstream dispersers, depending on their dispersal directions. We did not include salamanders that did not survive 10 weeks post-release in this comparison. Body condition was shown as the residual to the regression line of the cubed root of mass and total length constructed by all 31 salamanders before release [46]. We also compared sedentariness, mean daily movement, and LHR of salamanders at the Donghe River among seasons, using a Friedman rank sum test with a post-hoc Conover test. We defined seasons as: spring – March to May, summer – June to August, autumn - September to November, and winter - December to February. The two summers (2013 and 2014) were treated separately. We excluded salamanders that did not survive at least four seasons, as well as those that moved long distances downstream during floods. Only two salamanders at the Heihe River were eligible under these criteria. We did not conduct seasonal comparison for these two animals due to the small sample size, nor did we add them into the Donghe group for comparison since they were from a different system. All statistics were conducted in R (version 3.3.3, [47]), using the package PMCMR [48] and differences were considered significant at p < 0.05.

#### **Abbreviations**

CGS: Chinese giant salamander; IUCN: International Union for Conservation of Nature and Natural Resources; JGS: Japanese giant salamander; LHR: Linear home range;; MCP: Minimum convex polygons

#### Acknowledgements

We thank the two field assistants – Yan-hua Jiao at the Heihe River and Sheng-zhi Luo at the Donghe River for their continual monitoring work during the entire project period. We also thank Dr. Carrie Vance for editing assistance and Drs. Guiming Wang and Qing Zhao for their help on statistical analyses.

#### Authors' contributions

AJK, SW, and HXZ conceived the original idea and procured funding. LZ, AJK, and HXZ designed the study. LZ, HZ, QW, and WJ conducted field work and collected data. LZ and AJK analyzed data and wrote the manuscript, and all other authors provided editorial advice for the manuscript. All authors read and approved the final manuscript.

#### **Funding**

This work was funded by: 1) the U.S. Forest Service International Programs; 2) Ocean Park Conservation Foundation, Hong Kong; 3) the Shaanxi science and technology plan projects (2013KW20); 4) the Shaanxi Institute of Zoology; 5) Mississippi State University, and 6) Sun Yat-sen University (17lgpy113). The following zoos also contributed funds to support the project including: 1) Memphis Zoo; 2) Los Angeles Zoo; 3) Omaha Henry Doorly Zoo; and 4) Fort Worth Zoo. None of the funders had any input into the content of the manuscript.

#### Availability of data and materials

The datasets used and/or analyzed during the current study are available from the corresponding author on reasonable request.

#### Ethics approval and consent to participate

Ethical treatment, handling and tracking of giant salamanders were approved by the Shaanxi Institute of Zoology animal research committee. The Fisheries Bureau of Shaanxi Province provided a permit for reintroduction of giant salamanders at both streams.

#### Consent for publication

Not applicable.

#### Competing interests

The authors declare that they have no competing interests.

#### **Author details**

<sup>1</sup>School of Life Sciences, Sun Yat-sen University, 135 West Xingang Road, Guangzhou 510275, Guangdong Province, China. <sup>2</sup>Department of Wildlife, Fisheries and Aquaculture, Mississippi State University, Starkville, MS 39762, USA. <sup>3</sup>Biochemistry, Molecular Biology, Entomology, and Plant Pathology, Mississippi State University, Starkville, MS 39762, USA. <sup>4</sup>Shaanxi Institute of Zoology, 88 Xingqing Road, Xi'an 710032, Shaanxi Province, China. <sup>5</sup>Conservation and Research Department, Memphis Zoo, 2000 Prentiss Place, Memphis, TN 38112, USA. <sup>6</sup>Department of Wildlife, Fisheries and Aquaculture, Thompson Hall, Box 9690, Mississippi State University, Starkville, MS 39759, USA.

Zhang et al. BMC Zoology (2019) 4:7 Page 11 of 11

Received: 8 May 2019 Accepted: 16 September 2019 Published online: 21 October 2019

#### References

- 1. IUCN Red List version 2018–2. Available from: https://www.iucnredlist.org/.
- Young BE, Lips KR, Reaser JK, Ibáñez R, Salas AW, Cedeño JR, et al. Population declines and priorities for amphibian conservation in Latin America. Conserv Biol. 2001;15:1213–23.
- Stuart SN, Chanson JS, Cox NA, Young BE, Rodrigues AS, Fischman DL, et al. Status and trends of amphibian declines and extinctions worldwide. Science. 2004;306:1783–6.
- Pounds JA, Bustamante MR, Coloma LA, Consuegra JA, Fogden MP, Foster PN, et al. Widespread amphibian extinctions from epidemic disease driven by global warming. Nature. 2006;439:161.
- Fisher MC, Henk DA, Briggs CJ, Brownstein JS, Madoff LC, McCraw SL, et al. Emerging fungal threats to animal, plant and ecosystem health. Nature. 2012;484:186
- Lips KR, Reeve JD, Witters LR. Ecological traits predicting amphibian population declines in Central America. Conserv Biol. 2003;17:1078–88.
- Ripple WJ, Wolf C, Newsome TM, Hoffmann M, Wirsing AJ, McCauley DJ. Extinction risk is most acute for the world's largest and smallest vertebrates. Proc Natl Acad Sci. 2017;114:10678–83.
- Hammerson G., Phillips C. Cryptobranchus alleganiensis. The IUCN Red List of Threatened Species 2004: e.T59077A11879843. Available from: https://doi. org/10.2305/IUCN.UK.2004.RLTS.T59077A11879843.en.
- Kaneko Y., Matsui M. Andrias japonicus. The IUCN Red List of Threatened Species 2004: e.T1273A3376261. Available from: https://doi.org/10.2305/ IUCN.UK.2004.RLTS.T1273A3376261.en.
- Liang G, Geng B, Zhao E. Andrias davidianus. IUCN Red List Threat Species. 2004;e.T1272A3375181. Available from: https://doi.org/10.2305/IUCN.UK. 2004.RLTS.T1272A3375181.en.
- Germano JM, Bishop PJ. Suitability of amphibians and reptiles for translocation. Conserv Biol. 2009;23:7–15.
- Taguchi Y. Seasonal movements of the Japanese giant salamander (Andrias japonicus): evidence for possible breeding migration by this stream-dwelling amphibian. Jpn J Ecol Jpn. 2009;59:117–28.
- Zheng H, Wang X. Telemetric data reveals ecologically adaptive behavior of captive raised Chinese giant salamanders when reintroduced into their native habitat. Asian Herpetol Res. 2010;1:31–5.
- Burgmeier NG, Sutton TM, Williams RN. Spatial ecology of the eastern hellbender (*Cryptobranchus alleganiensis alleganiensis*) in Indiana. Herpetologica. 2011;67:135–45.
- Armstrong DP, Seddon PJ. Directions in reintroduction biology. Trends Ecol Evol. 2008;23:20–5.
- Knox CD, Monks JM. Penning prior to release decreases post-translocation dispersal of jewelled geckos. Anim Conserv. 2014;17:18–26.
- Nickerson MA, Mays CE. A study of the Ozark hellbender Cryptobranchus alleganiensis bishopi. Ecology. 1973;54:1164–5.
- Bodinof CM, Briggler JT, Junge RE, Beringer J, Wanner MD, Schuette CD, et al. Postrelease movements of captive-reared Ozark hellbenders (Cryptobranchus alleganiensis bishopi). Herpetologica. 2012;68:160–73.
- Larson KA, Gall BG, Briggler JT. The use of gastric transmitters to locate nests and study movement patterns of breeding male Ozark hellbenders (Cryptobranchus alleganiensis bishopi). Herpetol Rev. 2013;44:434–9.
- Wang X, Zhang K, Wang Z, Ding Y, Wu W, Huang S. The decline of the Chinese giant salamander *Andrias davidianus* and implications for its conservation. Oryx. 2004;38:197–202.
- Pan Y, Wei G, Cunningham AA, Li S, Chen S, Milner-Gulland EJ, et al. Using local ecological knowledge to assess the status of the critically endangered Chinese giant salamander *Andrias davidianus* in Guizhou Province, China. Oryx. 2016;50:257–64.
- Chen S, Cunningham AA, Wei G, Yang J, Liang Z, Wang J, et al. Determining threatened species distributions in the face of limited data: spatial conservation prioritization for the Chinese giant salamander (*Andrias davidianus*). Ecol Evol. 2018;8:3098–108.
- 23. Turvey ST, Chen S, Tapley B, Wei G, Xie F, Yan F, et al. Imminent extinction in the wild of the world's largest amphibian. Curr Biol. 2018;28:R592–4.
- Cogan CB, D'Elia J, Convery K, Brandt J, Bulgerin T. Analysis of California condor (*Gymnogyps californianus*) activity using satellite telemetry data. Open Ornithol J. 2012;5:82–93.

- Cunningham AA, Turvey ST, Zhou F, Meredith HM, Guan W, Liu X, et al. Development of the Chinese giant salamander *Andrias davidianus* farming industry in Shaanxi Province, China: conservation threats and opportunities. Oryx. 2016;50:265–73.
- Zhang L, Jiang W, Wang Q-J, Zhao H, Zhang H-X, Marcec RM, et al. Reintroduction and post-release survival of a living fossil: the Chinese giant salamander. PLoS One. 2016;11:e0156715.
- Zhang L, Wang Q, Willard ST, Jiang W, Zhang H, Zhao H, et al. Environmental characteristics associated with settlement of reintroduced Chinese giant salamanders. J Herpetol. 2017;51:417–24.
- Lowe WH. Linking dispersal to local population dynamics: a case study using a headwater salamander system. Ecology. 2003;84:2145–54.
- Cecala KK, Price SJ, Dorcas ME. Evaluating existing movement hypotheses in linear systems using larval stream salamanders. Can J Zool. 2009;87:292–8.
- Eggert C. Use of fluorescent pigments and implantable transmitters to track a fossorial toad (*Pelobates fuscus*). Herpetol J. 2002;12:69–74.
- Blomquist SM, Hunter ML Jr. A multi-scale assessment of habitat selection and movement patterns by northern leopard frogs (*Lithobates* [*Rana*] pipiens) in a managed forest. Herpetol Conserv Biol. 2009;4:142–60.
- Blais DP. Movement, home range, and other aspects of the biology of the eastern hellbender (*Cryptobranchus alleganiensis alleganiensis*): a radio telemetric study [M.S. Thesis]. Binghamton: State University of New York; 1996.
- Tuberville TD, Clark EE, Buhlmann KA, Gibbons JW. Translocation as a conservation tool: site fidelity and movement of repatriated gopher tortoises (*Gopherus polyphemus*). Anim Conserv. 2005;8:349–58.
- 34. Davis JM, Stamps JA. The effect of natal experience on habitat preferences. Trends Ecol Evol. 2004;19:411–6.
- 35. Ousterhout BH, Luhring TM, Semlitsch RD. No evidence of natal habitat preference induction in juveniles with complex life histories. Anim Behav. 2014;93:237–42.
- Chen Y, Yang A, Wang W, Wu Z, Bai H. Influence of water temperature and body mass on food intake of Chinese giant salamanders. J Hydroecology. 2006;32:47.
- Ultsch GR. Ecology and physiology of hibernation and overwintering among freshwater fishes, turtles, and snakes. Biol Rev. 1989;64:435–515.
- Storey KB, Storey JM. Natural freeze tolerance in ectothermic vertebrates. Annu Rev Physiol. 1992;54:619–37.
- Preatoni D, Mustoni A, Martinoli A, Carlini E, Chiarenzi B, Chiozzini S, et al. Conservation of brown bear in the Alps: space use and settlement behavior of reintroduced bears. Acta Oecol. 2005;28:189–97.
- Grabowski TB, Jennings CA. Post-release movements and habitat use of robust redhorse transplanted to the Ocmulgee River. Georgia Aquat Conserv Mar Freshw Ecosyst. 2009;19:170–7.
- Weise FJ, Lemeris JR Jr, Munro SJ, Bowden A, Venter C, van Vuuren M, et al. Cheetahs (*Acinonyx jubatus*) running the gauntlet: an evaluation of translocations into free-range environments in Namibia. PeerJ. 2015;3:e1346.
- Banks PB, Norrdahl K, Korpimäki E. Mobility decisions and the predation risks of reintroduction. Biol Conserv. 2002;103:133–8.
- 43. Pittman SE, Osbourn MS, Semlitsch RD. Movement ecology of amphibians: a missing component for understanding population declines. Biol Conserv.
- Semlitsch RD. Differentiating migration and dispersal processes for pondbreeding amphibians. J Wildl Manag. 2008;72:260–7.
- Marcec R, Kouba A, Zhang L, Zhang H, Wang Q, Zhao H, et al. Surgical implantation of coelomic radiotransmitters and postoperative survival of Chinese giant salamanders (*Andrias davidianus*) following reintroduction. J Zoo Wildl Med. 2016;47:187–95.
- Wheeler BA, Prosen E, Mathis A, Wilkinson RF. Population declines of a longlived salamander: a 20+–year study of hellbenders, *Cryptobranchus alleganiensis*. Biol Conserv. 2003;109:151–6.
- R Core Team. R: a language and environment for statistical computing. Vienna: R: foundation for statistical Computing; 2016.
- Pohlert T. The pairwise multiple comparison of mean ranks package (PMCMR). 2014. Available from: http://CRAN.R-project.org/package=PMCMR.

#### **Publisher's Note**

Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.